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Arboreal Urban Cooling Is Driven by Leaf Area Index, Leaf Boundary Layer Resistance, and Dry Leaf Mass per Leaf Area: Evidence from a System Dynamics Model

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Abstract: Heat waves are becoming more frequent due to climate change. Summer heat waves can be particularly deadly in cities, where temperatures are already inflated by abundant impervious, dark surfaces (i.e., the heat island effect). Urban heat waves might be ameliorated by planting and maintaining urban forests. Previous observational research has suggested that conifers may be particularly effective in cooling cities. However, the observational nature of these studies has prevented the identification of the direct and indirect mechanisms that drive this differential cooling. Here, we develop a systems dynamics representation of urban forests to model the effects of the percentage cover of either conifers or broadleaf trees on temperature. Our model includes physiological and morphological differences between conifers and broadleaf trees, and physical feedback among temperature and energy fluxes. We apply the model to a case study of Vancouver, BC, Canada. Our model suggests that in temperate rainforest cities, conifers may be 1.0 °C cooler than broadleaf trees; this differential increases to 1.2 °C when percentage tree cover increases from 17% to 22% and to 1.7 °C at 30% cover. Our model suggests that these differences are due to three key tree traits: leaf area index, leaf boundary layer resistance, and dry mass per leaf area. Creating urban forests that optimize these three variables may not only sequester CO₂ to mitigate global climate change but also be most effective at locally minimizing deadly urban heat waves.

Keywords: heat wave; urban heat island; climate adaptation; microclimate; conifer; local climate; urban planning; human health; broadleaf trees; system dynamics model



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1. Introduction

Heat waves are becoming more frequent due to climate change [1]. More frequent heat waves are especially harmful in cities because temperatures are already inflated by abundant impervious, dark surfaces (i.e., the heat island effect); air quality may often be low; and there are many vulnerable city residents [2–5]. High urban temperatures can lead to direct human and non-human animal deaths [6]. In summer 2021, a heat wave associated with anthropogenic climate change [7] on the Pacific coast of North America broke Canada's temperature record and was associated with quadrupled human mortality [8]. Heat waves can also cause indirect effects [9], such as wildfires [10]. As climate change intensifies, ameliorating urban heat waves is becoming urgent [11,12].

The intensity of urban heat waves might be lessened by planting and maintaining urban forests [13–15]. Trees cool cities in multiple ways, including via evapotranspiration [16] and by blocking solar radiation from reaching the ground [17]. In addition to cooling locally, trees may also cool globally as part of efforts to sequester carbon dioxide [18,19]. Trees may be key to simultaneously tackling heatwaves at local and global scales.

Previous observational research has suggested that conifers may be particularly effective in cooling cities when compared with broadleaf trees in temperate [20] and Mediterranean climates [21]. This increased cooling capacity may be due to a variety of tree

characteristics, such as leaf area index and evapotranspiration rate [21–24]. Yet the observational nature of these studies hampers identification of the key arboreal characteristics that drive cooling in cities.

Physical urban temperature models could help identify the key cooling characteristics of trees. Yet despite the documented variability across tree types, urban temperature models have not yet accounted for this variability in tree physiognomy (e.g., [25–28]).

Here, we leverage the modeling tool Stella to build a systems dynamics model of urban temperature associated with different percentage cover of either conifers or broadleaf trees. Our model includes measured physiological and morphological differences between conifers and broadleaf trees and physical feedbacks among temperature and heat fluxes. Our objective is to use this model to identify the tree traits that create differences in urban cooling.

We analyze the model in a case study of Vancouver, BC, Canada, which both has experienced high heat wave mortality [6,8] and has plans to increase its tree cover from 17% to 22% by 2050 [29]. Currently, approximately 22% of the city's urban forest is coniferous, based on data from Metro Vancouver [30] (see Figure 1). Vancouver is a large city (>600,000 residents) in the extreme southwest of Canada with warm wet winters and moderate dry summers. Streets are forested primarily by non-native broadleaf deciduous trees [31], while native forests consist mostly of conifers, including Douglas fir (*Pseudotsuga menziesii*), western hemlock (*Tsuga heterophylla*), and western red cedar (*Thuja plicata* [32]). Studies have shown a large heat island effect in the city (up to 11.6 °C; [3]), as well as a park cooling effect of 1–5 °C extending outward from city parks [33]. These effects are most important in the summer, when high temperatures can lead to mortality. We expect that urban temperatures will be lower when forested with conifers and that this effect will increase as the forest cover increases. Moreover, we expect that a few conifer traits will drive this temperature difference. This study may help inform what trees Vancouver and other cities can plant and maintain in order to ameliorate heat waves.

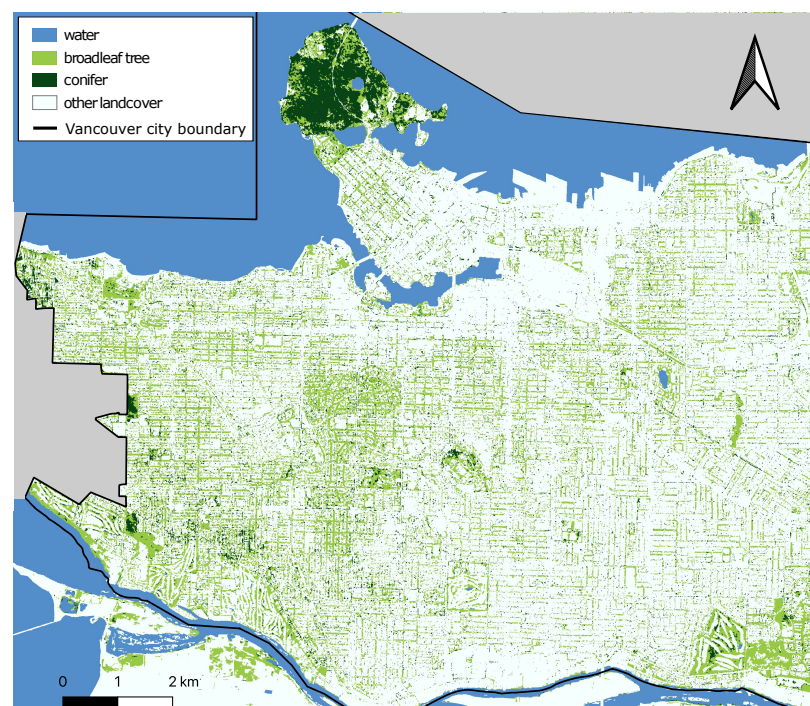


Figure 1. Map of the city of Vancouver in extreme southwestern British Columbia, Canada, showing 2019 forest cover of conifers and broadleaf trees. Note that most trees within the urban matrix are broadleaf trees. Land cover data from Metro Vancouver [30].

2. Materials and Methods

2.1. System Dynamics Model

We built a system dynamics model of the heat exchange and feedback among trees and the city using Stella Architect version 2.1.5 (2629) (ISEE Systems, Lebanon, NH, USA). Stella Architect is a widely used system dynamics modeling tool that can capture the multiple feedback loops within complex physical systems [34,35]. We included heat fluxes from solar radiation, longwave radiation, conduction, convection, and evapotranspiration among air, ground, built environment, and trees (Figure 2). Rural air is assumed to be unaffected by the urban heat island and such cool air interfacing with urban air is key way that natural landscapes cool cities. We treated the built environment and tree heating mechanisms in separate modules. The model was run separately with two sets of parameters: one set of parameters represented evergreen, needle-bearing conifers (signified below by “(conifer)”) and the other broadleaf, deciduous angiosperms (signified below by “(broadleaf)”). This structure enabled comparison of heat exchange and temperature between conifers and broadleaf trees. We included variables and interdependencies that are likely to affect tree and city temperatures [36,37].

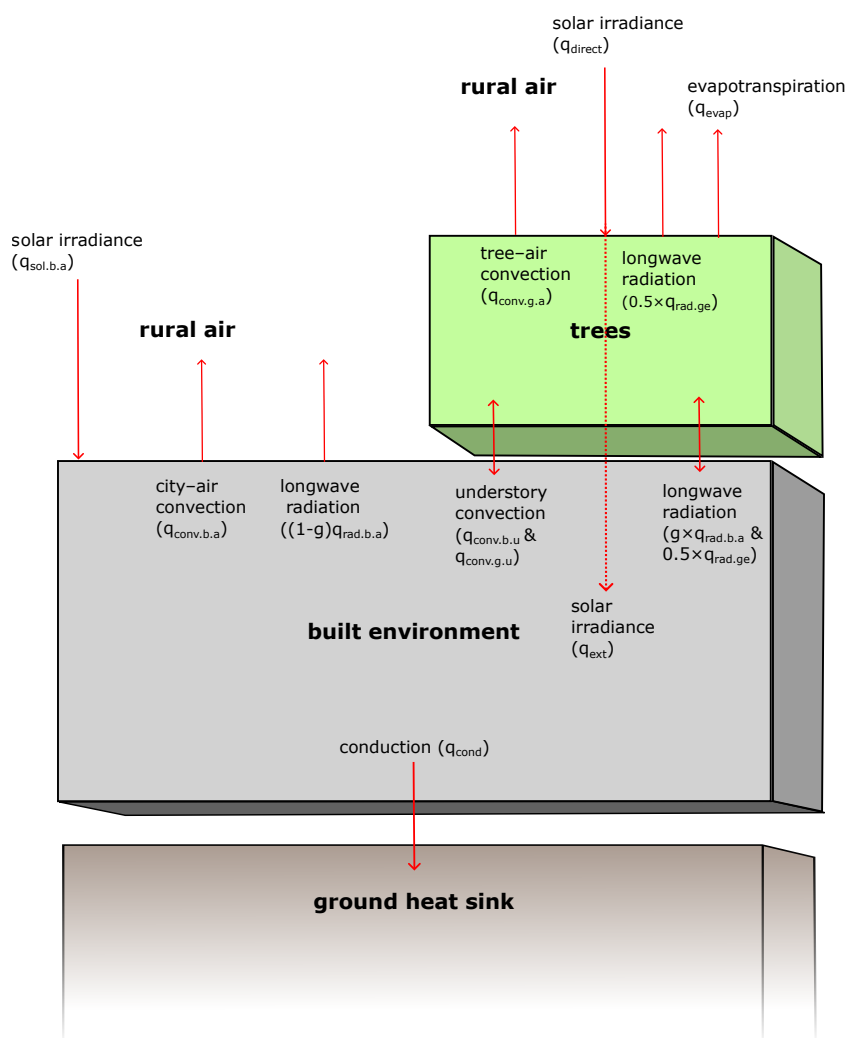


Figure 2. Model structure showing heat fluxes among city, trees, rural air, and ground.

All calculations were carried out on a per-area basis. The modeled timeframe was set to 6 d (144 h) in order to provide sufficient time to overcome the initial temperature condition. All visualizations and model inference was based on the last 24 h of the timeframe (i.e., 120–144 h). Rather than setting the sum of fluxes equal to zero and solving for the equilibrium temperature [25], we solved for the dynamic, instantaneous temperature using

differential equations. We used Euler integration and a time step of 1 min ($DT = 1/60$) to solve the differential equations. To verify that this numerical simulation was robust, we tested an integration time of 1 s ($DT = 1/3600$) and the Runge–Kutta 4 (RK4) integration method and achieved identical results. After creating the model, we conducted a sensitivity analysis, varying the tree parameters to identify which conifer–broadleaf differences were most important in driving differences in temperature.

Parameter values were set to values representative of 1 August in Vancouver, BC (Table 1), including the current urban forest cover of 17%, target 2050 forest cover of 22% [29], and 30%. The first author lived for five years in this region, and their first-hand experiences of how temperature varied near different urban trees motivated the conceptualization of this study.

Table 1. Parameter values used in the model.

Parameter	Value	Unit	Source
T_{built}	init. = 22	°C	
C_v	3×10^6	J/m ³ /K	[38]
s	3600	s/h	Number of seconds in 1 h
H	0.3	m	Optimized to obtain reasonable city temperatures
g	17, 22, or 30	dimensionless [29]	
α_b	0.114	dimensionless	Average albedo in Vancouver, BC based on albedo values from [20] and percentage landcover values from the Land Cover Classification 2014—5 m Hybrid (updated in November 2019) dataset from http://www.metrovancouver.org/data (accessed on 26 January 2021) for all landcover classes except deciduous and coniferous trees, water, shadow, clouds, and ice
t_s	5.75	h	Time of sunrise on August 1 in Vancouver, BC, Canada
L	$49.2827 \frac{2\pi}{360}$	Radians	Latitude of Vancouver, BC
D	213	day	Day of year for 1 August 2020
ϵ_b	0.9	dimensionless	[38]
σ	5.67×10^{-8}	W/m ² /K ⁴	
v	4.54	m/s	Average of hourly wind velocities recorded at Vancouver Airport on 1 August 2020, via https://weatherspark.com/h/d/476/2020/8/1/Historical-Weather-on-Saturday-August-1-2020-in-Vancouver-Canada#Figures-WindSpeed (accessed on 18 January 2023)
$T_{air.low}$	12.2	°C	Historical August 1 low in rural Bowen Island, BC. Data from https://www.accuweather.com/en/ca/bowen-island/v0n/july-weather/53179 (accessed on 18 January 2023)
$T_{air.high}$	20.5	°C	Historical August 1 high in rural Bowen Island, BC. Data from https://www.accuweather.com/en/ca/bowen-island/v0n/july-weather/53179 (accessed on 18 January 2023)
k_g	1.65	W/m/K	Following [25]
T_g	8	°C	Following [25]
d	2	m	Following [25]
T_{tree}	init. = 22	°C	
m_a (conifer)	0.263	kg/m ²	[39,40]
m_a (broadleaf)	0.073	kg/m ²	[39,40]
f_w	0.7	dimensionless	[39,41]
c_{water}	4188	J/kg/K	
c_{dry}	1396	J/kg/K	[39]
H_r	72.7	Average humidity on 1 August 2020 in Vancouver, BC	

Table 1. Cont.

Parameter	Value	Unit	Source
ϵ_c	0.046	dimensionless	[42]
K (conifer)	0.52	dimensionless	[43]
K (broadleaf)	0.7	dimensionless	[25]
ϵ_g	0.98	dimensionless	[44]
α_g (conifer)	0.08	dimensionless	[20]
α_g (broadleaf)	0.12	dimensionless	[20]
h_v	2450	J/kg	
M_w	0.018	kg/mol	
LAI (conifer)	8.6	dimensionless	LAI for <i>Pseudotsuga menziesii</i> [45]
LAI (broadleaf)	4.9	dimensionless	LAI for an oak–hickory forest [46]
$r_{cuticular.abaxial}$ (conifer)	30,303	s/m	For <i>Pseudotsuga menziesii</i> [36,47]
$r_{cuticular.abaxial}$ (broadleaf)	8500	s/m	For <i>Acer platanoides</i> [36,48]
$r_{intracellular.abaxial}$	17.5	s/m	Average for typical plant [36,49]
$r_{stomatal.abaxial}$ (conifer)	140	s/m	For <i>Abies lasiocarpa</i> [36,50]
$r_{cuticular.adaxial}$ (conifer)	30,303	s/m	For <i>Pseudotsuga menziesii</i> [36,47]
$r_{cuticular.adaxial}$ (broadleaf)	8500	s/m	For <i>Acer platanoides</i> [36,48]
$r_{intracellular.adaxial}$	17.5	s/m	Average for typical leaf [36,49]
$r_{stomatal.adaxial}$ (conifer)	1×10^6	s/m	<i>Pseudotsuga menziesii</i> lacks stomata on the adaxial surface [51]
$r_{stomatal.adaxial}$ (broadleaf)	1×10^6	s/m	Most broadleaf trees lack stomata on the adaxial surface [39]
c_p	29.2	J/mol/K	[39]

2.1.1. Module 1: Built Environment Model

The temperature of the built environment, T_{built} , was initialized at 22 °C. Change in T_{built} is driven by heat fluxes into and out of the built environment:

$$\Delta T_{built} C = q_{sol.b} + q_{rad.b} - q_{conv.b} - q_{cond}, \quad (1)$$

where C is the heat capacity of the built environment (J/m²/K), $q_{sol.b}$ is the solar heat flux entering the built environment (J/m²/h), $q_{rad.b}$ is the net longwave radiation entering the built environment (J/m²/h), $q_{conv.b}$ is the net convection heat flux exiting the built environment (J/m²/h), and q_{cond} is the net conduction heat flux exiting from the built environment into the ground (J/m²/h). The ground is assumed to be a uniform dirt-like material that is representative of the average underlying substrate across the city.

$$C = C_v H, \quad (2)$$

where C_v is the volumetric heat capacity of the built environment (J/m³/K) [38] and H is its effective height (m).

$$q_{sol.b} = q_{ext} + q_{sol.b.a}, \quad (3)$$

$$q_{sol.b.a} = (1 - \alpha_b)(1 - g)snI, \quad (4)$$

where q_{ext} (J/m²/h) is the solar irradiance that goes through the tree canopy to the built environment below (see Equation (19)); $q_{sol.b.a}$ is the solar irradiance directly entering the built environment; α_b is the albedo of the built environment, i.e., the fraction of solar radiation reflected by the built environment (dimensionless); g is the fraction of the built environment covered by trees (dimensionless); s is the number of seconds in an hour (to convert the quantity from J/s to J/h); and I is the direct normal irradiance (DNI; J/m² normal to the sun per second) taken from the Physical Solar Model (PSM3) of the National Solar Radiation Database (NSRDB; <https://nsrdb.nrel.gov/data-viewer>; accessed on 24 May 2022) [52] for 1 August 2020 in Vancouver, BC (location ID = 262013). We chose this date because it is, on average, the hottest day of the year (<https://weatherspark.com/m/476/8/Average-Weather-in-August-in-Vancouver-Canada>; accessed on 22 January 2021). The same direct normal irradiance values are used for each of the six model run days. n is

the proportion of the direct normal irradiance that is incident onto a square meter tangent to the earth (dimensionless) and is approximated by

$$n = \sin\left(\frac{1}{3.82}(t - t_s)\right) \cos(L - \delta), \quad (5)$$

where t is time (h) since the start of the model run (set to midnight) and t_s is the time of sunrise (h). L is the latitude and δ is the declination angle. All angles are measured in radians. δ is given by

$$\delta = 23.45 \frac{2\pi}{360} \sin\left(2\pi \frac{284 + D}{365}\right), \quad (6)$$

where D is the day of year (d).

$$q_{rad.b} = \frac{q_{rad.ge}}{2} - q_{rad.b.a}, \quad (7)$$

$$q_{rad.b.a} = \epsilon_b \sigma T_{built}^4, \quad (8)$$

where $q_{tree.ge}$ is the longwave radiation emitted by the trees (J/m²/h; divided by two because only half of the emitted radiation reaches the built surface). ϵ_b is the emissivity of the built environment, and σ is the Stefan–Boltzmann constant (W/m²/K⁴).

$$q_{conv.b} = q_{conv.b.a} + q_{conv.b.u}, \quad (9)$$

where $q_{conv.b.a}$ is the convection heat flux from the built environment to the rural air and $q_{conv.b.u}$ is the convection heat flux between the built environment and the air beneath the trees.

$$q_{conv.b.a} = s(1 - g)h(T_{built} - T_{air}), \quad (10)$$

where h is the urban convective heat transfer coefficient (J/K/s) and T_{air} is the temperature of the rural air (°C) and serves as the heat sink above the modeled region [26].

Following Silva et al. [26] and Rowley and Eckley [53],

$$h = 11.8 + 4.2v, \quad (11)$$

where v is the wind velocity.

$$T_{air} = \frac{T_{air.high} - T_{air.low}}{2} \sin\left(\frac{1}{3.82}(t - t_s) - \frac{\pi}{2}\right) + \frac{T_{air.high} + T_{air.low}}{2} \quad (12)$$

approximates the daily oscillation between the low rural air temperature ($T_{air.low}$, °C; experienced at sunrise) and the high rural air temperature ($T_{air.high}$, °C; experienced in the afternoon). All angles are in radians.

$$q_{conv.b.u} = gh(T_{built} - T_{understory}). \quad (13)$$

Because air at the built–tree interface would be similar to the tree temperature, $T_{understory}$ (°C) is approximated by the leaf temperature, T_{tree} (°C).

$$q_{cond} = \frac{k_g s (T_{built} - T_g)}{d}, \quad (14)$$

where k_g is the thermal conductivity of the ground (W/m/K), d is the ground depth (m), and T_g is the temperature (°C) of the ground at depth d and serves as the heat sink below the modeled region, following Pace et al. [25].

2.1.2. Module 2: Tree Model

The temperature of the urban forest, T_{tree} was initialized at 22 °C. Change in T_{tree} was driven by heat fluxes into and out of the tree canopy:

$$\Delta T_{tree} C_g g = q_{sol.g} - q_{evap} + q_{rad.g} - q_{conv.g}, \tag{15}$$

where C_g is the heat capacity of leaves (J/m²/K), $q_{sol.g}$ is the solar heat flux (J/m²/h), q_{evap} is the heat loss through evapotranspiration (J/m²/h), $q_{rad.g}$ is the net longwave radiation (J/m²/h), and $q_{conv.g}$ is the net convection heat flux (J/m²/h) [25,26].

$$C_g = \frac{m_a LAI}{1 - f_w} \left((1 - f_w) c_{dry} + (f_w c_{water}) \right), \tag{16}$$

where, following Equations (10) and (11) in [39], m_a is the dry leaf mass per unit leaf area (kg/m²), LAI is the leaf area index, f_w is the fraction of the leaf that is water (dimensionless), c_{dry} is the specific heat of dry leaf biomass (J/kg/K), and c_{water} is the specific heat of water (J/kg/K).

$$q_{sol.g} = q_{direct} - q_{ext}, \tag{17}$$

where q_{direct} is the sunlight energy that enters the trees (J/m²/h) and q_{ext} is the sunlight energy that passes through the canopy and enters the built environment beneath.

$$q_{direct} = LAI \times (1 - \alpha_g) g s n I (1 - \epsilon_c), \tag{18}$$

where α_g is the albedo of the leaves (dimensionless) and ϵ_c is the photosynthetic efficiency of the leaves [42].

$$q_{ext} = q_{direct} e^{-K \times LAI}, \tag{19}$$

where K is the extinction coefficient (dimensionless) according to the Beer–Lambert law [43].

$$q_{rad.g} = g q_{rad.b.a} - q_{rad.ge}, \tag{20}$$

$$q_{rad.ge} = 2 \times LAI \times \epsilon_g \sigma_s g T_{tree}^4, \tag{21}$$

where $q_{rad.b.a}$ is given in Equation (8) and ϵ_g is the emissivity of the leaves (dimensionless).

$$q_{evap} = E h_v g, \tag{22}$$

where E is the evapotranspiration rate (kg/m²/h) and h_v is the latent heat of vaporization of water (J/kg).

$$E = \frac{C_{leaf} - C_{air}}{r_{tot}} LAI \times s, \tag{23}$$

where C_{leaf} is the water vapor concentration at the evaporating surface within the leaf (kg/m³), C_{air} is the water vapor concentration in the air (kg/m³), and r_{tot} is the total resistance (s/m) [36].

Equation (23) assumes that higher leaf area index causes an increase in total evapotranspiration, which has been widely demonstrated in the literature [37,54]. Although there may be slightly lower per-leaf evapotranspiration at higher LAI values [37], we use a simple linear relationship. Other studies have mistakenly applied a per-leaf decrease in

evapotranspiration to the whole tree and treated LAI and E as inversely proportional for entire trees [25,55].

$$C_{leaf} = \frac{M_w e_{sat.leaf}}{RT_{tree}}, \tag{24}$$

$$C_{air} = \frac{M_w e_{air}}{RT_{air}}, \tag{25}$$

where M_w is the molecular weight of water (kg/mol), $e_{sat.leaf}$ is the saturation vapor pressure (Pa) of the leaf, e_{air} is the vapor pressure of the air, and R is the universal gas constant (J/mol/K).

$$e_{air} = \frac{H_r}{100} e_{sat.air}, \tag{26}$$

where H_r is the relative humidity (in percent) and $e_{sat.air}$ is the saturation vapor pressure of the air (Pa).

Following the World Meteorological Guide [56], the saturated vapor pressures of the leaf and air are approximated as follows:

$$e_{sat.leaf} = 100 \times 6.112 \exp\left(\frac{17.62T_{tree}}{T_{tree} + 243.12}\right), \tag{27}$$

$$e_{sat.air} = 100 \times 6.112 \exp\left(\frac{17.62T_{air}}{T_{air} + 243.12}\right). \tag{28}$$

The total resistance to evapotranspiration, r_{tot} in Equation (23), consists of resistance through the abaxial (lower) surface of the leaf ($r_{tot.abaxial}$; s/m) and the upper (adaxial) surface of the leaf ($r_{tot.adaxial}$; s/m), summed in parallel [36,39]:

$$r_{tot} = \frac{r_{tot.abaxial} r_{tot.adaxial}}{r_{tot.abaxial} + r_{tot.adaxial}}. \tag{29}$$

Because some resistances act in parallel and others in series, care must be taken in calculating the totals [39]. For the abaxial surface, the stomatal ($r_{stomatal.abaxial}$; s/m) and intracellular resistances ($r_{intracellular.abaxial}$; s/m) are added in series, and the resultant quantity is added in parallel with the cuticular resistance ($r_{cuticular.abaxial}$; s/m) [36]. This resultant quantity is then added in series with the boundary-layer resistance ($r_{air.abaxial}$; s/m) [36]:

$$r_{tot.abaxial} = \frac{1}{\frac{1}{r_{cuticular.abaxial}} + \frac{1}{r_{stomatal.abaxial} + r_{intracellular.abaxial}}} + r_{air.abaxial}. \tag{30}$$

The same is repeated for the adaxial surface:

$$r_{tot.adaxial} = \frac{1}{\frac{1}{r_{cuticular.adaxial}} + \frac{1}{r_{stomatal.adaxial} + r_{intracellular.adaxial}}} + r_{air.adaxial}. \tag{31}$$

Wind decreases the boundary layer resistance more for thin needles than larger broad leaves [36]:

$$r_{air.adaxial} = r_{air.abaxial}, \tag{32}$$

$$r_{air.abaxial}[\text{conifer}] = \frac{1000}{5.0 + 74.4v}, \tag{33}$$

$$r_{air.abaxial}[\text{broadleaf}] = \frac{100}{0.52 + 3.2v}, \tag{34}$$

where the conifer Equation (33) is derived from observations of *Abies amabilis* [57] and the broadleaf Equation (34) is derived from experiments with a synthetic broad leaf [58]. Studies have shown that even at night, tree stomata often remain open and evapotranspiration remains linked to water vapor deficit and wind speed [36], so we treat stomatal resistance as temporally constant for simplicity.

$$q_{conv.g} = q_{conv.g.a} + q_{conv.g.u} \quad (35)$$

where $q_{conv.g.a}$ is the convection heat flux (J/h) from trees to rural air and $q_{conv.g.u}$ is the convection heat flux from the bottom layer of leaves to the air above the built environment.

$$q_{conv.g.a} = 2c_p g_s (T_{tree} - T_{air}) \frac{LAI}{r_{air.abaxial}}, \quad (36)$$

where c_p is the specific heat of moist air at constant pressure (J/mol/K) [39]. Because the boundary resistance is equivalent for both sides of the leaf, we use just abaxial resistance in the equation for simplicity, and double the entire convection term, following [39].

$$q_{conv.g.u} = c_p g_s (T_{tree} - T_{built.undertree}) \frac{1}{r_{air.abaxial}}. \quad (37)$$

Because air at the built environment beneath the tree is likely similar to the built environment temperature, $T_{built.undertree}$ (°C) is approximated by the built environment temperature, T_{built} (°C). We assume that only the lower surface of the bottom layer of leaves interacts with the undertree air.

3. Results

Our model results were consistent with empirical observations of tree temperature—our modeled maximum broadleaf and conifer temperatures were within 1 °C of those temperatures measured in Vancouver, BC, based on Landsat 8 imagery [20].

At 17% tree cover, the maximum temperature of the built environment was 1.0 °C warmer when it was forested by conifers than by broadleaf trees (Figure 3a). When tree cover increased to 22%, the temperature differential increased to 1.2 °C (Figure 3b); when tree cover increased to 30%, conifers were 1.7 °C cooler (Figure 3c).

Solar radiation and convection dominated heat fluxes into and out of the built environment (Figure 4). Solar radiation was slightly higher in the broadleaf-forested built environment (because $q_{ext}[\text{broadleaf}] > q_{ext}[\text{conifer}]$), and the convection heat flux was higher at night (Figure 4). Conduction also played an important role, but longwave radiation was negligible (Figure 4).

Similarly, solar radiation and convection dominated the heat fluxes into and out of the urban forests, though convection curves were shaped quite differently across the urban forest types (Figure 5). Broadleaf forests showed a single large diurnal flux of convection, while coniferous forests also showed a smaller diurnal flux and a substantial nocturnal flux. Evapotranspiration was significant in the coniferous forest, but small in the broadleaf forest (Figure 6). Longwave radiation was insignificant in both forest types (Figure 5).

Sensitivity analysis revealed that three variables drive the conifer–broadleaf temperature differential. Setting the conifer values of leaf area index (LAI), leaf boundary layer resistance ($r_{air.adaxial}$ and $r_{air.abaxial}$), and dry leaf mass per leaf area (m_a) equal to the respective broadleaf values for these three variables removed differences between conifer and broadleaf temperatures, both for the urban forest and for the built environment (Figure 7). That is, controlling for these three variables produces equivalent temperatures for both conifer-forested cities and broadleaf-forested cities (Figure 7).

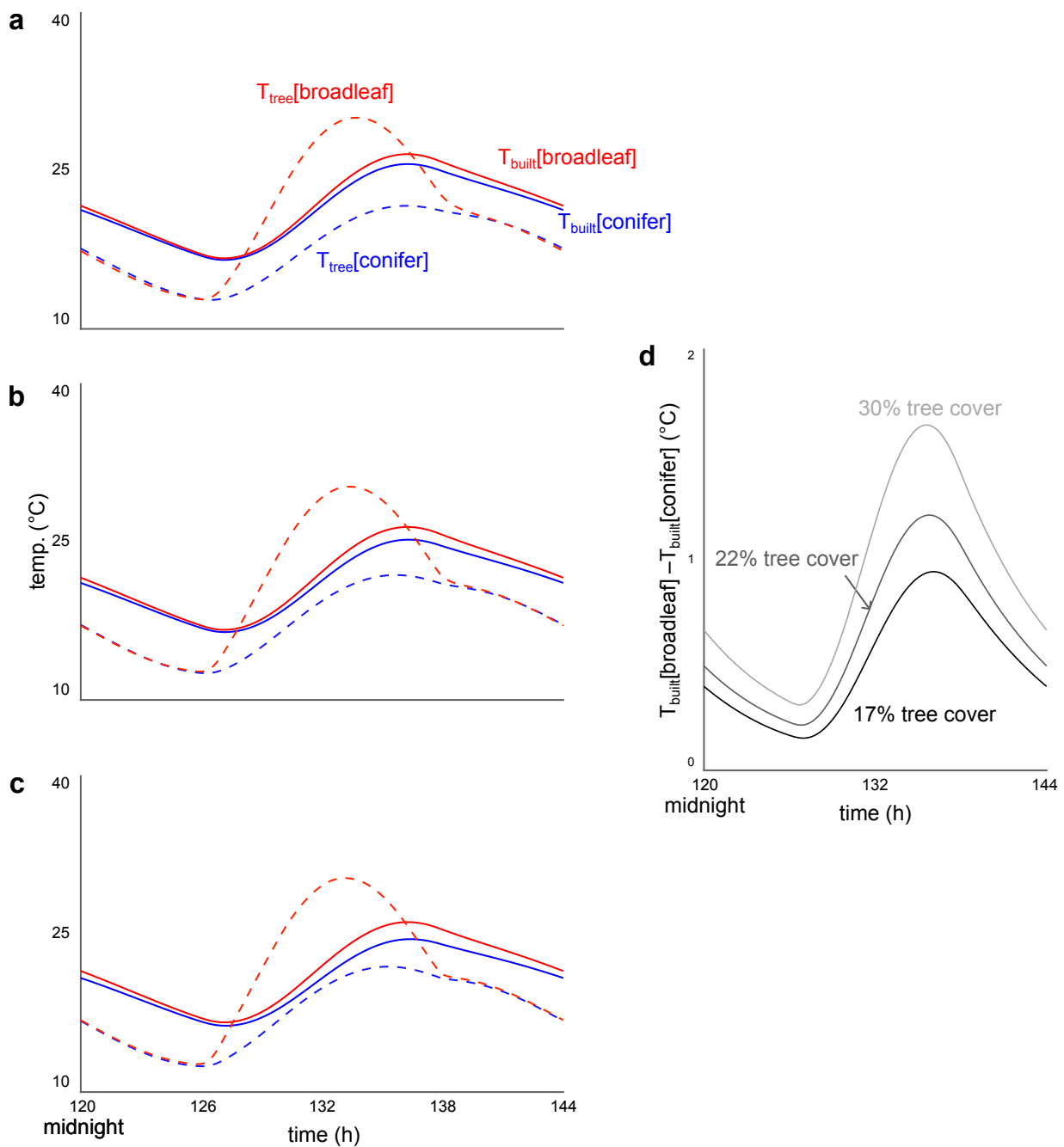


Figure 3. Temperatures of the built environment (T_{built} , solid lines) and urban forests (T_{tree} , dashed lines) in the conifer (blue) and broadleaf tree models when (a) 17% of the city is covered by trees, (b) 22% of the city is covered by trees, and (c) 30% of the city is covered by trees. (d) shows the built environment temperature differential when forested by different tree types at 17%, 22%, and 30% tree cover.

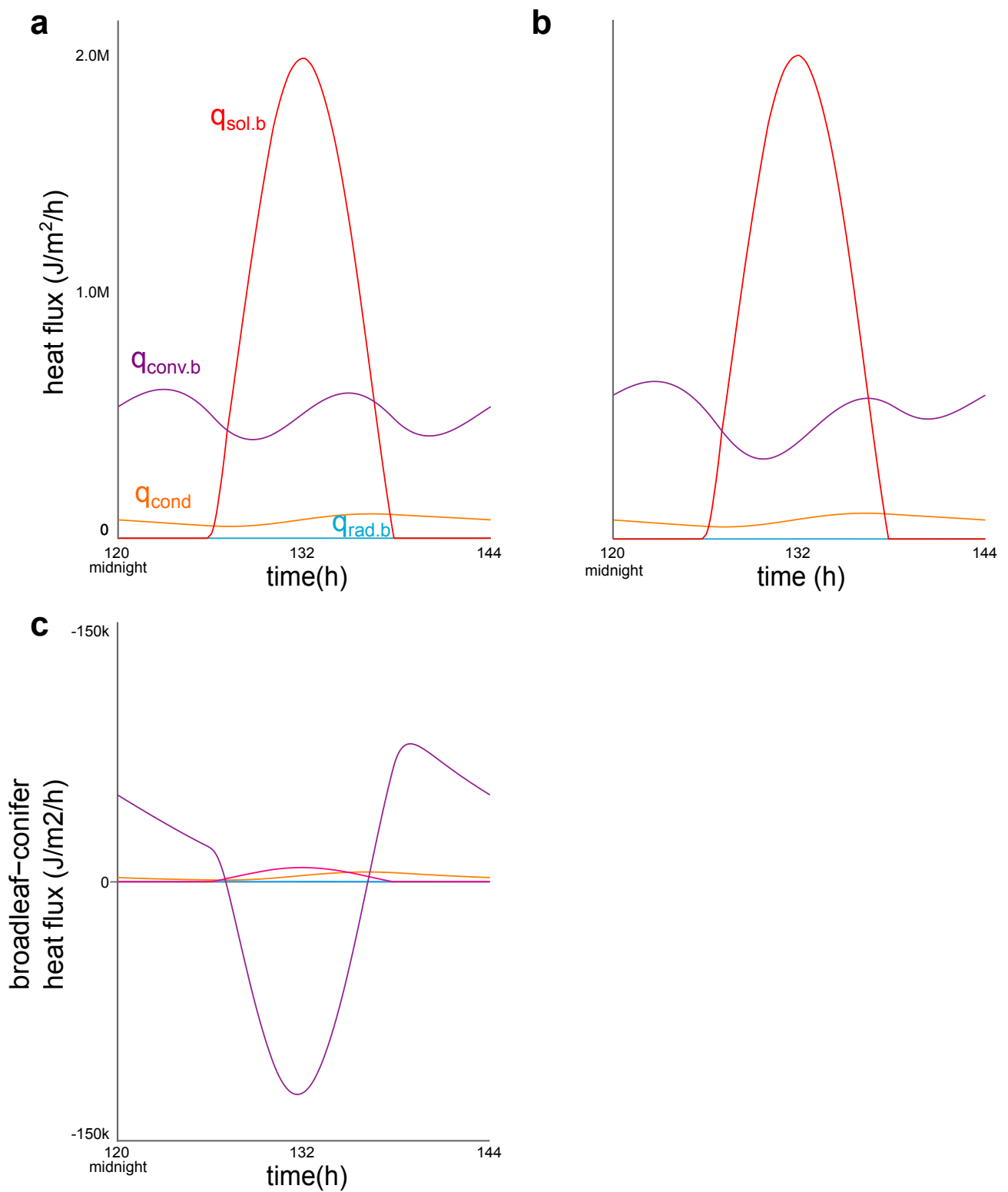


Figure 4. Heat fluxes into and out of the built environment when 17% forested by (a) conifers or (b) broadleaf trees. (c) shows the differences between these fluxes by forest type.

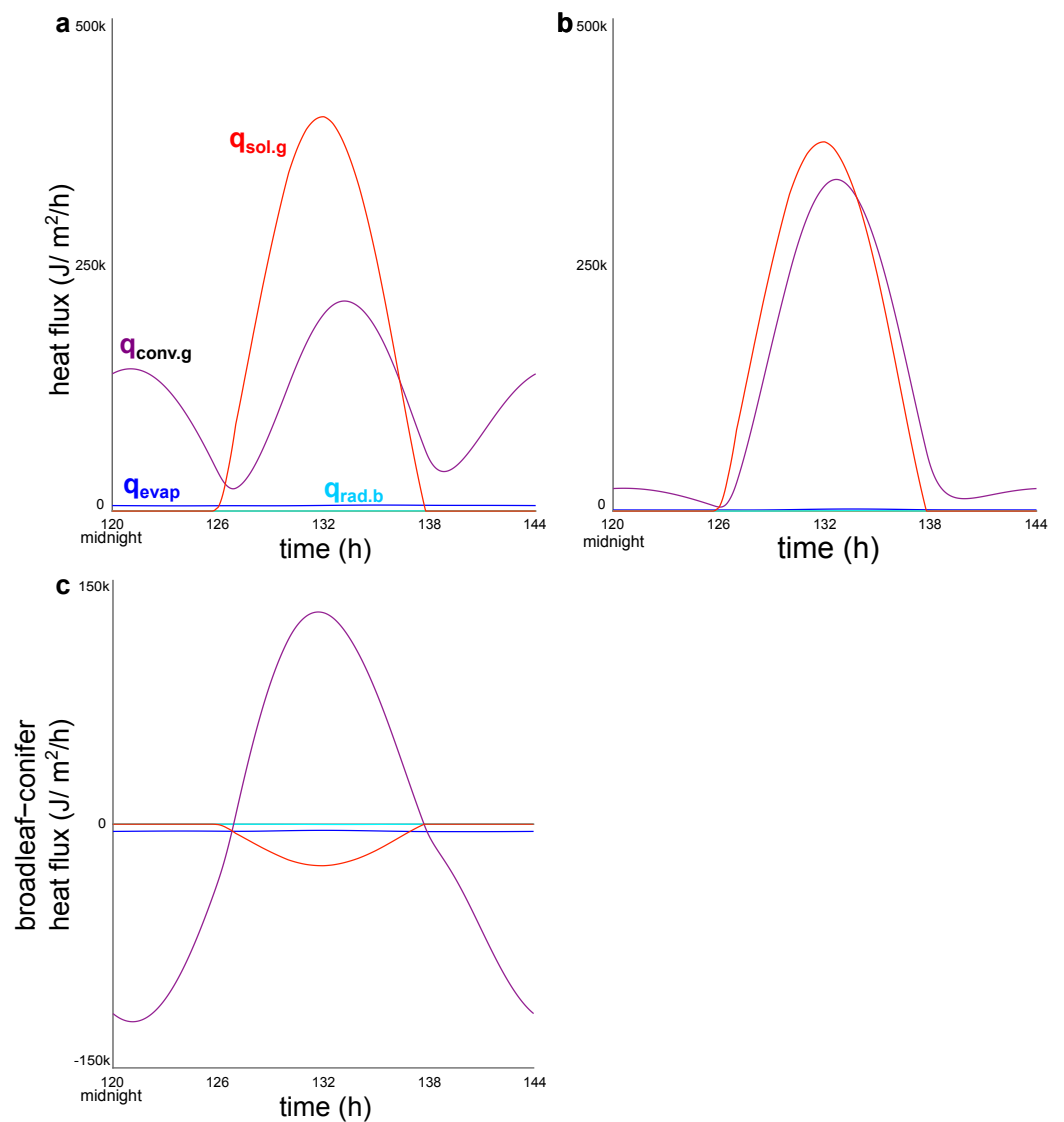


Figure 5. Heat fluxes into and out of the urban forest of (a) conifer or (b) broadleaf trees. (c) shows the differences between these fluxes by forest type. Urban forest covers 17% of city.

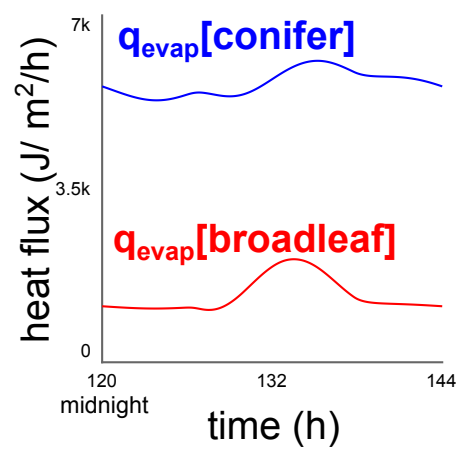


Figure 6. Evapotranspiration heat flux for conifers and broadleaf urban forests, where urban forest covers 17% of the city.

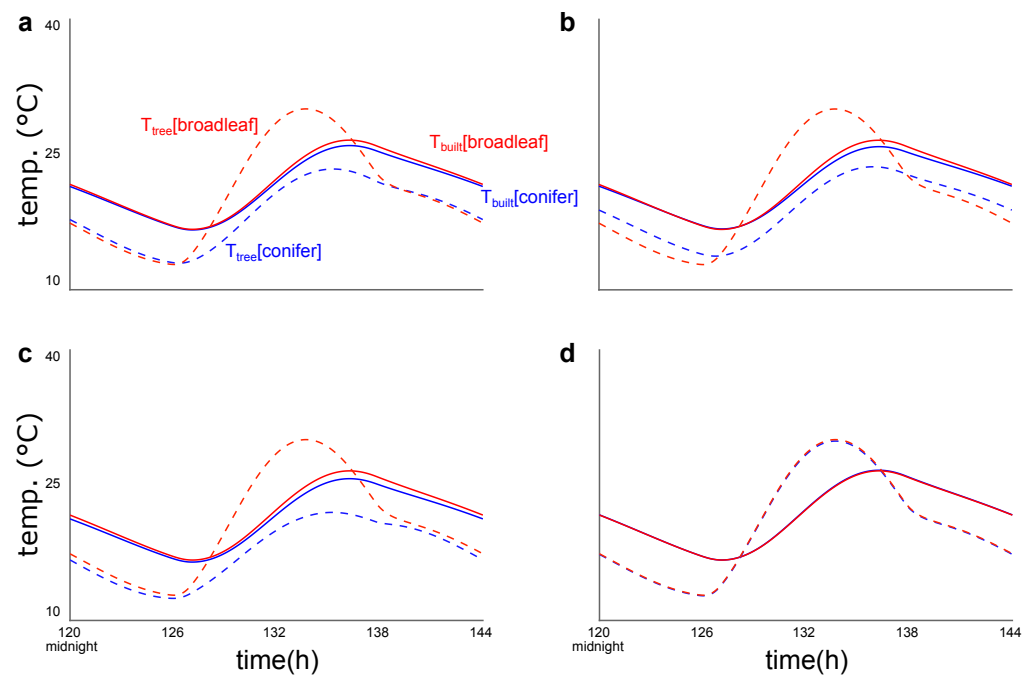


Figure 7. Sensitivity analysis results showing built environment temperatures (solid lines) and tree temperatures (dashed lines) when (a) the conifer leaf area index (LAI) is set equal to that of broadleaf trees ($LAI(\text{conifer}) = LAI(\text{broadleaf})$), (b) conifer leaf boundary layer resistances are set equal to those of broadleaf leaves ($r_{air.adaxial}(\text{conifer}) = r_{air.adaxial}(\text{broadleaf})$ and $r_{air.abaxial}(\text{conifer}) = r_{air.abaxial}(\text{broadleaf})$), (c) conifer dry leaf mass per unit leaf area was set equal to that of broadleaf trees ($m_a(\text{conifer}) = m_a(\text{broadleaf})$), and (d) when conifer values for all three variables (leaf area index, boundary layer resistance, dry leaf mass) were set equal to the broadleaf values. All plots are for 17% tree cover.

4. Discussion

This study harnessed a physical systems dynamics model of urban temperature and conifer and broadleaf tree traits to identify the degree to which conifers cool cities compared to broadleaf trees, and which key traits drive these differences. We found that conifers cool cities more than broadleaf trees and that this effect increases at higher tree cover (Figure 3). Our model suggests that this differential cooling can be fully explained by conifer–broadleaf differences in terms of three traits: leaf area index, leaf boundary layer resistance, and leaf dry mass per leaf area (Figure 7). Relative to broadleaf values, conifer values for these traits yield higher evapotranspiration, lower built environment solar radiation, and higher daytime convection, which in turn produce cooler tree and built environment temperatures (Figure 7).

Our results support the City of Vancouver’s plan to increase forest cover to 22% and its emphasis on planning native conifers such as Douglas fir (*Pseudotsuga menziesii*) [29], not only because they support birds [59] and mammals and help lessen flooding [60], but because they may be more effective at cooling the city.

Our finding that leaf area index is important for urban cooling is consistent with other studies (e.g., [61–65]). However, our leaf area index results and methods are inconsistent with physical models that assumed that single-leaf relationships between leaf area index and evapotranspiration can be applied to entire trees [25,55]. Future work might further clarify the degree to which the effect of leaf area index on evapotranspiration decays at high leaf area index values.

Our results are also consistent with a smaller body of literature on the role of boundary layer resistance in increasing evapotranspiration and lowering leaf temperature (e.g., [66,67]). Studies have often focused on stomatal resistance (or conductance) (e.g., [65,67]), but we found that boundary layer resistance was more important in cooling.

Our findings showing a negative relationship between leaf mass per leaf area and temperature are inconsistent with some studies that have shown that thick leaves are hotter [67,68]. However, the thicker leaves in these studies were likely also wider and exhibited higher leaf boundary layer resistances and lower surface-area-to-volume ratios. Thus, the particular combination of traits exhibited simultaneously by conifers appear to be uniquely important.

Beyond the conifer–broadleaf tree dichotomy, selectively planting trees that holistically exhibit high leaf area index, low leaf boundary layer resistance, and high dry leaf mass per leaf area may help cities most effectively combat increasingly frequent heat waves. Nonetheless, the cooling effect is only one of many factors that are important when considering which trees belong in a particular part of a particular city [69,70].

The three-fold greater evapotranspiration exhibited by conifers (Figure 6) is consistent with observational evidence [37], and confirms hypotheses that heightened evapotranspiration may lower conifer foliage surface temperature [20]. Moreover, the higher diurnal broadleaf tree convection suggests that, relative to conifers, broadleaf trees may warm pedestrians more and offer less respite from heat waves. When choosing to reforest or afforest cities, tree type matters.

Although strategies to mitigate high urban temperatures often focus on increasing surface albedo (e.g., [26,71]), our results suggest that strategies that increase albedo may sometimes create hotter temperatures. Specifically, our model showed that while conifers exhibit lower albedo, their higher leaf area index means that they also block more light from passing through the canopy to enter and warm the built environment below (Equation (23), Figure 4). We second Yang et al. [72] in cautioning city planners to not treat albedo as a “silver bullet”.

Studies have shown that cooling by trees varies in response to wind, moisture content, latitude, and other climate variables [15,21,33,73]. Our model is parameterized for climates, cities, and trees of the Pacific Northwest, and so our results are not necessarily indicative of conifers or broadleaf trees per se, but of the relationships among urban forests and the Pacific Northwest urban ecosystem [74]. However, our model could be extended to other regions by replacing the values from Table 1 with those relevant for a given location, and perhaps making some variables dynamic that we treated as static, and vice versa.

We included arboreal variables that are suggested to have large effects on tree temperatures. These variables proved sufficient to reproduce empirical tree temperature observations [20]. However, incorporating additional variables and feedbacks may improve the model [36]. For example, feedbacks among photosynthesis rates, CO₂ concentration, water availability, and stomatal aperture may be important [36]. Future work might test the importance of these variables and relationships.

The spatial locations of trees matter, not just for cooling [21,24], but also for equitably serving urban residents [29,70,75]. Future models might account for spatial relationships among trees and the built environment, relationships with urban non-human animals, and geographic patterns and processes of systemic racism and oppression that have left cities inequitably forested [76].

We follow [27] in making all our model code and parameter values publicly available (cf. [25,26,28]). We hope this will help other researchers build upon our model and add complexity that we have overlooked. We invite other urban climate scholars to join us in making code and parameter values public.

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